

Aqueous oxidation of isoprene-derived organic aerosol species as a source of atmospheric formic and acetic acids

Received 00th January 20xx,
Accepted 00th January 20xx

Kelvin H. Bates,^{*a,b,†} Daniel J. Jacob,^b James D. Cope,^a Xin Chen,^{c,*} Dylan B. Millet,^d and Tran B. Nguyen^a

DOI: 10.1039/x0xx00000x

Atmospheric chemistry models generally assume organic aerosol (OA) to be photochemically inert. Recent mechanisms for the oxidation of biogenic isoprene, a major source of secondary organic aerosol (iSOA), produce excessive OA in the absence of subsequent OA reactivity. At the same time, models underestimate atmospheric concentrations of formic and acetic acids for which OA degradation could provide a source. Here we show that the aqueous photooxidation of an isoprene-derived organosulfate (2-methyltriolsulfate or MTS), an important iSOA component, produces formic and acetic acids in high yields and at timescales competitive with deposition. Experimental data are well fit by a kinetic model in which three sequential oxidation reactions of the isoprene organosulfate produce two molar equivalents of formic acid and one of acetic acid. We incorporate this chemistry and that of 2-methyltetrol, another ubiquitous iSOA component, into the GEOS-Chem global atmospheric chemistry model. Simulations show that photooxidation and subsequent revolatilization of this iSOA may account for up to half of total iSOA loss globally, producing 4 Tg a⁻¹ each of formic and acetic acids. This reduces model biases in gas-phase formic acid and total organic aerosol over the Southeast United States in summer by ~30% and 60% respectively. While our study shows the importance of adding iSOA photochemical sinks into atmospheric models, uncertainties remain that warrant further study. In particular, improved understanding of reaction dependencies on particle characteristics and concentrations of particle-phase OH and other oxidants are needed to better simulate the effects of this chemistry on the atmospheric budgets of organic acids and iSOA.

Introduction

Formic and acetic acids are ubiquitous and abundant in earth's troposphere, and play important roles in determining the acidity of cloudwater and rainfall.¹⁻⁶ By affecting the acidity of aerosol particles as well, they influence cloud nucleation⁷ and the aqueous chemistry of other aerosol constituents.⁸⁻⁹ Formic and acetic acids are directly emitted to the atmosphere by plants,¹⁰⁻¹² fires,¹³⁻¹⁵ soils,¹⁶⁻¹⁸ ice,¹⁹ and anthropogenic activities,²⁰⁻²³ and can also be produced in the atmosphere through the gas-phase chemistry of biogenic,^{5,24-29} pyrogenic,³⁰⁻³³ and anthropogenic^{23,34-36} organic compounds. However,

atmospheric modeling studies have repeatedly shown that these known sources are insufficient to account for the observed abundances of formic and acetic acids, especially in areas with strong biogenic influence.^{5,37-40}

One route to organic acid formation not generally included in models is the aqueous chemistry of biogenic organic compounds in aerosol and cloudwater (Figure 1).⁴¹⁻⁴⁷ Isoprene is an important primary precursor to biogenic secondary organic aerosol (SOA), especially via the intermediate formation of isoprene epoxydiols (IEPOX).⁴⁸⁻⁴⁹ Experimental evidence has shown that IEPOX efficiently produces 2-methyltetrols (1,2,3,4-tetrahydroxy-2-methylbutane; 2-MT) and their sulfate esters (2-MTS) following its uptake into sulfate-containing deliquesced particles.⁴⁸⁻⁵¹ These compounds have been observed as significant fractions of organic aerosol in biogenically influenced regions around the world.⁵²⁻⁵⁶ Many models treat isoprene-derived SOA (iSOA) as inert, with deposition as its only loss, which may explain why the most up-to-date model mechanisms tend to overestimate its atmospheric abundance.⁵⁷⁻⁵⁸ Photochemical sinks of aqueous SOA, such as particle-phase oxidation reactions followed by revolatilization of fragmentation products, have been suggested as a possible mechanism to help balance the SOA budget.⁵⁹⁻⁶² If these fragmentation products include formic and acetic acids, such photochemical sinks of SOA could simultaneously help correct the model underestimates for these acids.

^a Department of Environmental Toxicology, University of California Davis, Davis, CA 95616, United States

^b School of Engineering and Applied Sciences, Harvard University, Cambridge, MA 02138, United States

^c Department of Civil and Environmental Engineering, Massachusetts Institute of Technology, Cambridge, MA 02139, United States

^d Department of Soil, Water, and Climate, University of Minnesota, St. Paul, MN 55108, United States

[†] Now at: NOAA Chemical Sciences Laboratory, Earth System Research Laboratories, and Cooperative Institute for Research in Environmental Sciences, University of Colorado, Boulder, CO 80305, United States

*Now at: Department of Civil and Environmental Engineering, Massachusetts Institute of Technology, Cambridge, MA 02139, United States

Electronic Supplementary Information (ESI) available: additional figures, tables, and mechanisms related to experimental and modelling results. See DOI: 10.1039/x0xx00000x

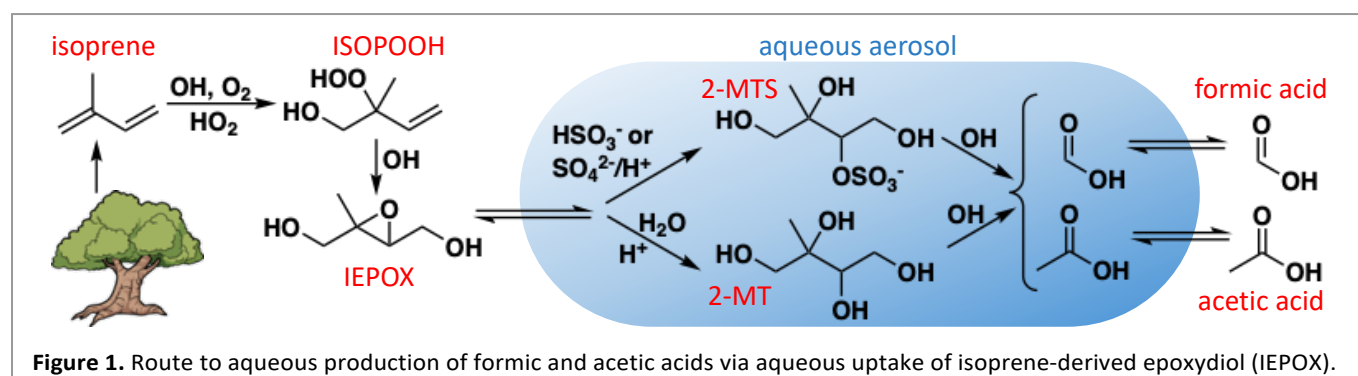


Figure 1. Route to aqueous production of formic and acetic acids via aqueous uptake of isoprene-derived epoxydiol (IEPOX).

We recently showed using nuclear magnetic resonance (NMR) spectroscopy that the aqueous oxidation of 2-MT by OH produces large yields of formic and acetic acids.⁶³ Here we extend that analysis to a synthetic 2-MTS isomer with similar experimental techniques and use a kinetic model to constrain organic acid yields from its oxidation. We then implement the aqueous reactions of 2-MT and 2-MTS into GEOS-Chem, a global chemical transport model, to assess the contributions of these pathways both to the global iSOA and organic acid budgets.

Experimental

Photochemical experiments. NMR experiments to determine formic and acetic acid yields from the aqueous photochemistry of 2-MTS followed the procedures described by Cope *et al.*⁶³ for the study of 2-MT. In each of four experiments, 80–400 mM of 2-MTS was mixed with 0.8–4 M H₂O₂ (50% in H₂O, Sigma Aldrich) in D₂O, and the solution was irradiated directly in the quartz NMR tube using a UV-B broadband fluorescent light (peak emission at 310 nm; photon flux shown in Figure S1) within a photochemical enclosure. At regular intervals, the tube was removed from the photochemical enclosure and ¹H NMR data were collected on a 400 MHz Bruker Avance III HD instrument (Figure S2). Data were analyzed using TopSpin, with suppression of the H₂O peak using standard WATERSUP parameters. Control experiments without H₂O₂ showed no significant loss of 2-MTS.

Experiments were performed with two methods to test the dependence of measured organic acid yields on potential gas-liquid partitioning, as in Cope *et al.*⁶³ In the first, cyclohexane in CDCl₃ was placed in a narrow glass capillary, which was flame sealed and dropped into the NMR tube containing 2-MTS and H₂O₂ in D₂O prior to the experiment to act as an internal standard. In the second, cyclohexane in CDCl₃ was instead placed in the NMR tube itself, and the 2-MTS reaction mixture was sealed in a glass capillary and dropped into the tube. In the former experiments, the 2-MTS and H₂O₂ solution is exposed to air in the headspace of the NMR tube during the reaction and products can partition between phases, while in the latter experiments the glass capillary contains only an aqueous phase.

2-methyl-1,2,3-trihydroxy-4-butylsulfate, the isomer of 2-MTS investigated in this work, was synthesized according to previously published methods.^{64,65} All reagents were purchased from Sigma Aldrich. Briefly, 2-methylvinylloxirane (95%) was converted to 1,2-dihydroxy isoprene (DHI) by acid-catalyzed hydrolysis, which was then epoxidized with 3-chloroperbenzoic

acid (≤77%) in acetonitrile at 0° C to yield d1-IEPOX. Reaction of d1-IEPOX with tetrabutylammonium hydrogen sulfate (97%) then afforded 2-methyl-1,2,3-trihydroxy-4-butylsulfate, the purity of which (>95%) was confirmed by ¹H NMR.

Kinetic modelling. To constrain the organic acid yields and multigenerational aqueous chemistry of 2-MTS, we use a simple kinetic model (Mechanism S1) run on Matlab (MathWorks, Inc.). The mechanism uses light flux as measured in the photochemical reaction chamber (Figure S1) to initiate OH radical production via photolysis of H₂O₂, and includes the reactions OH + H₂O₂, HO₂ + HO₂, and OH + HO₂.⁶⁶ We also include the aqueous reactions of formic and acetic acid with OH,⁶⁷ and use the OH + 2-MTS oxidation rate constant ($k_{\text{MTS}+\text{OH}} = 1.52 \times 10^9 \text{ M}^{-1} \text{ s}^{-1}$) recently measured by Abellar *et al.*⁶⁵ for the 2-MTS isomer used here. We initialize the model with aqueous concentrations of 2-MTS and H₂O₂ used in each experiment and vary the direct organic acid yields from 2-MTS, the number and OH oxidation rates of intermediate stable products of 2-MTS + OH, and the organic acid yields from those stable products to find the best fit to measured organic acid yields.

Global modelling. We use the atmospheric chemistry model GEOS-Chem (<http://geos-chem.org>) to simulate the aqueous chemistry of 2-MT and 2-MTS on a global scale. GEOS-Chem is driven by assimilated meteorological observations from the NASA Modern-Era Retrospective analysis for Research and Applications (MERRA-2). We use model version 12.8 (DOI: 10.5281/zenodo.3860693) with complex SOA chemistry⁶⁸ as a base and perform global simulations at 2°×2.5° horizontal resolution with 47 vertical layers. Regional results reported for the Amazon (53.75 to 76.25°W, 11°S to 3°N), Southeast United States (81.25 to 93.75°W, 31 to 39°N), and Congo Basin (11.25 to 28.75°E, 5°S to 5°N) are from the same global simulations. Transport and convection are calculated at 15 min intervals, including advection based on the TPCORE algorithm,⁶⁹ nonlocal boundary layer mixing,⁷⁰ and convective mass fluxes from MERRA-2. Emissions and chemistry are calculated at intervals of 30 min. We use the standard Harmonized Emission Component (HEMCO) configuration in GEOS-Chem for emissions,⁷¹ which includes biogenic emissions calculated from the Model of Emissions of Gases and Aerosols from Nature (MEGAN) version 2.1 inventory with isoprene emissions scaled to 535 Tg a⁻¹,⁷² anthropogenic emissions from the Community Emissions Data System (CEDS) inventory,⁷³ and biomass burning emissions from

the Global Fire Emissions Database (GFED4s).⁷⁴ CEDS emissions in Africa are overwritten with the Diffuse and Inefficient Combustion Emissions inventory (DICE-Africa).⁷⁵

GEOS-Chem uses a recently compiled mechanism of isoprene chemistry⁵⁸ based on the Reduced Caltech Isoprene Mechanism described by Wennberg *et al.*⁷⁶ The model also includes the reactive uptake of IEPOX into aqueous aerosol as implemented by Marais *et al.*,⁷⁷ with an aqueous-phase first order rate coefficient dependent on particle acidity and on the concentrations of sulfate, bisulfate, and nitrate ions ($k_{aq} = k_1[H^+] + k_2[SO_4^{2-}][H^+] + k_3[HSO_4^-] + k_4[NO_3^-][H^+]$).^{78,79} In the base model, IEPOX reactive uptake produces a single inert SOA species (SOAIE); here, we separate the reactive uptake into one acid-dependent reaction producing 2-MT ($k_1 = 3.6 \times 10^{-2} \text{ M}^{-1} \text{ s}^{-1}$), one sulfate- and bi-sulfate-dependent reaction producing 2-MTS ($k_2 = 2.0 \times 10^{-4} \text{ M}^{-2} \text{ s}^{-1}$, $k_3 = 7.3 \times 10^{-4} \text{ M}^{-1} \text{ s}^{-1}$), and one nitrate-dependent reaction producing 2-methyltetrol nitrates (2-MTN; $k_4 = 2.0 \times 10^{-4} \text{ M}^{-2} \text{ s}^{-1}$), where the individual rate coefficients k_{1-4} are identical to those used in the base model. We then add 2-MT + OH aqueous chemistry under acidic conditions from Cope *et al.*⁶³ and 2-MTS + OH aqueous chemistry from this study (Mechanism S1) to the GEOS-Chem mechanism. All formic and acetic acid produced from 2-MT and 2-MTS chemistry is allowed to partition instantaneously back to the gas phase. 2-MTN, which accounts for <1% of IEPOX uptake in our simulations, is still treated as inert. For aqueous OH, we use the parameterization described by Jacob *et al.*⁸⁰ in which $[OH]_{aq} = (1 \times 10^{-19} \text{ M cm}^3 \text{ molecule}^{-1}) [OH]_g$.

We further update the “base” GEOS-Chem mechanism with additional formic acid sources as described by Chen *et al.*,⁸¹ building on the prior work of Paulot *et al.*⁵ and Millet *et al.*³⁹ These updates include phototautomerization of acetaldehyde,⁸² oceanic emissions of alkene precursors to secondary formic acid,^{5,39} agricultural and soil emissions of formic acid scaled to ammonia and nitrogen oxide fluxes,⁵ and uptake of formic acid on ice and dust.⁵ We also include the newly implemented gas-phase chemistry of DHI as described by Bates *et al.*,⁸³ which incorporates DHI formation from isoprene hydroxynitrate hydrolysis,⁸⁴ and the reaction of isoprene hydroxy-hydroperoxide with aqueous bisulfite.^{85,86} All results are for annual or seasonal simulations using meteorology from 1 December 2015 to 30 November 2016, following a six-month model initialization.

Results and Discussion

Photochemical experiments and kinetic models. Figure 2 shows the combined results of the aqueous-phase NMR experiments in which 2-MTS was oxidized by OH along with results from the kinetic model. In order to compare results across experiments, the reaction time coordinate was converted to reaction extent (equivalent to $k_{\text{MTS}+\text{OH}} \times [\text{OH}] \times \text{time}$) by taking the natural log of the inverse of the fraction of 2-MTS remaining in the NMR tube. When 2-MTS concentrations were too low to quantify (<10 mM), reaction extent was extrapolated as simple exponential decay from previous points. The kinetic model could then be fit to the combined data from all

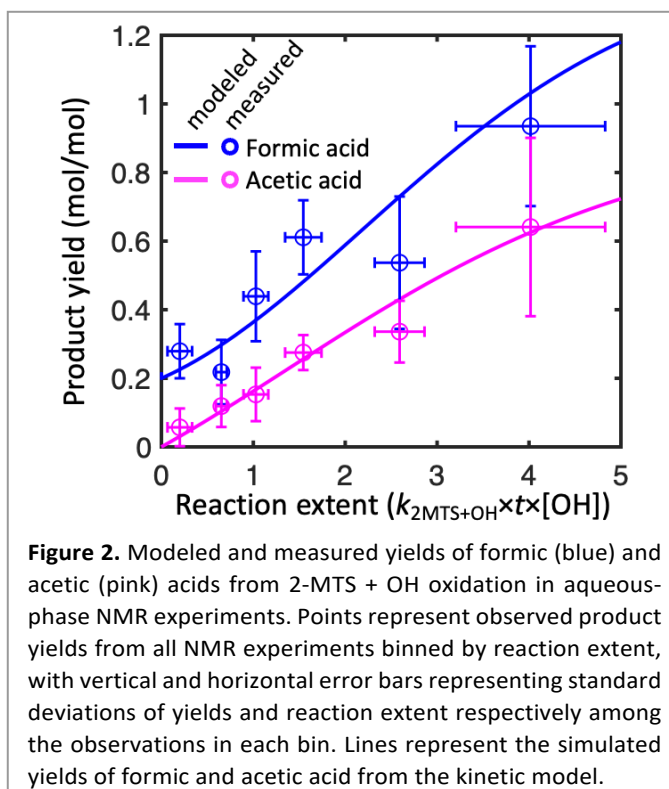


Figure 2. Modeled and measured yields of formic (blue) and acetic (pink) acids from 2-MTS + OH oxidation in aqueous-phase NMR experiments. Points represent observed product yields from all NMR experiments binned by reaction extent, with vertical and horizontal error bars representing standard deviations of yields and reaction extent respectively among the observations in each bin. Lines represent the simulated yields of formic and acetic acid from the kinetic model.

experiments. Measurements are binned by reaction extent in Figure 2 for ease of interpretation; full plots of each experiment are shown in Figure S3. The kinetic model satisfactorily captures the acid yield trends and timing in individual experiments (Figure S3) in addition to the binned experimental results as a function of reaction extent.

Observed yields (moles of acid in solution per mole of 2-MTS reacted) of formic and acetic acid from 2-MTS oxidation are high, generally exceeding 100% and 30% respectively by the time 2-MTS is below detection limits. Actual yields are higher owing to the simultaneous oxidative losses of formic and acetic acids during the experiment. Yields are not constant with time but rather rise sharply with reaction extent, implying that the acids are produced in multigenerational oxidation processes. Production of formic acid starts promptly upon irradiation of the solution, with an initial yield of 20–35%, while observed acetic acid yields start near zero, suggesting that formic acid is made both in the initial and later oxidation steps while acetic acid is only produced in a multistep reaction mechanism.

In experiments in which the 2-MTS and H_2O_2 solution was sealed in the capillary, no remaining 2-MTS was detected after 2–3 hours; in contrast, experiments with the 2-MTS and H_2O_2 solution outside the capillary required up to 8 hours to reach a point at which 50% of the 2-MTS had been oxidized. For experiments in the former category, because the capillary is sealed without headspace, organic acids cannot partition to the gas phase, which likely causes differences in pH between sealed and unsealed experiments. Other differences between experiments include initial 2-MTS concentrations, distance between the NMR tube and light source, and attenuation of some UV-B light by the glass capillary in the quartz NMR tube. Despite differences in experimental design, no statistically

significant difference in yields as a function of effective OH exposure ($[\text{OH}] \times \text{time}$) was observed between the two experimental methods, so we aggregate results from all experiments together in our analysis.

To best fit experimental yields while accounting for simultaneous oxidative losses of the organic acids, the 2-MTS oxidation mechanism used in the kinetic model (Mechanism S1) requires overall yields of 2 moles of formic acid and 1 mole of acetic acid per mole of 2-MTS oxidized, implying that four of the five carbon atoms in 2-MTS are converted to small organic acids during its aqueous oxidation. In the best-fit mechanism, these yields are spread over three oxidative steps, which is necessary to account for the prompt production of formic acid and then allow both yields to rise as the reaction progresses. The optimized fit also requires that the second and third oxidation steps progress a factor of two more slowly than the initial reaction of 2-MTS with OH. The kinetic mechanism for 2-MTS is similar to that proposed for 2-MT by Cope *et al.*⁶³ (1.5 moles of formic acid and 1.25 moles of acetic acid over three oxidative steps per mole of 2-MT reacted), but with a higher formic:acetic acid ratio and faster later-generation oxidation reactions.

Figure 3 shows the temporal evolution of aqueous atmospheric 2-MTS and 2-MT chemistry in the optimized mechanism under representative conditions as implemented in GEOS-Chem. To properly track aerosol mass and oxidation state in the model, we assign chemical formulas to the oxidative intermediates of both precursors; these formulas (see Mechanism S1) are not meant to represent actual species, but rather the approximate elemental ratios of aqueous products remaining at each reaction step. Based on the small fraction of mass lost to formic and acetic acids in the first oxidation step and the likelihood that the reaction adds oxygen to the molecule, we assume that the first-generation intermediates from both 2-MT and 2-MTS are slightly more massive than their parent compounds, resulting in an initial increase in both molar mass and oxidation state. We assume that the final carbon from

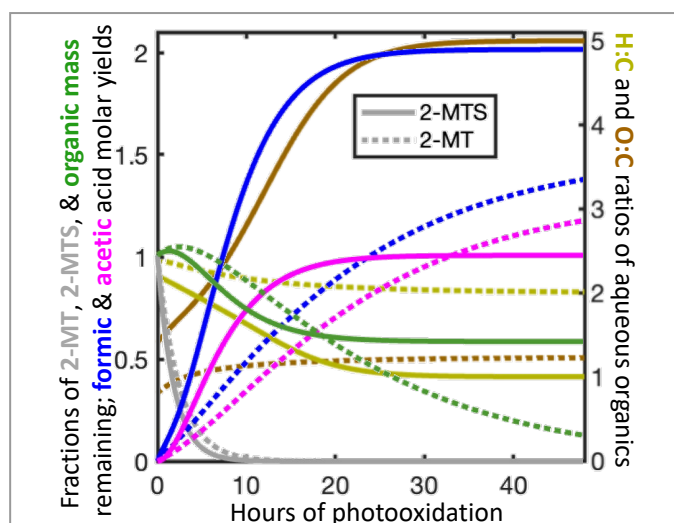


Figure 3. Temporal evolution of aqueous atmospheric chemistry of the 2-MTS (solid lines) and 2-MT (dashed lines) systems in the optimized mechanism (Mechanism S1) as implemented in GEOS-Chem for $[\text{OH}]_g = 10^6 \text{ molecules cm}^{-3}$.

2-MTS that does not form formic or acetic acid remains in the particle phase as a C_1 organosulfate, resulting in a high particulate O:C ratio of 5 and a final organic mass 58% as large as the initial mass. In contrast, the remaining carbon from 2-MT is presumed to form CO_2 , removing all organic mass from the particle phase.⁶³ The optimized mechanism is meant as a simple parameterization of the reaction progress for modeling; for example, it cannot rule out the possibility that a larger number of faster oxidation steps are involved in the overall reaction process, and the chemical formulas assigned to intermediates are not constrained by our experimental results.

Global modeling. Figure 4 shows the change in annually averaged surface IEPOX-derived secondary organic aerosol (IEPOX-SOA) loadings when the reactions from the kinetic

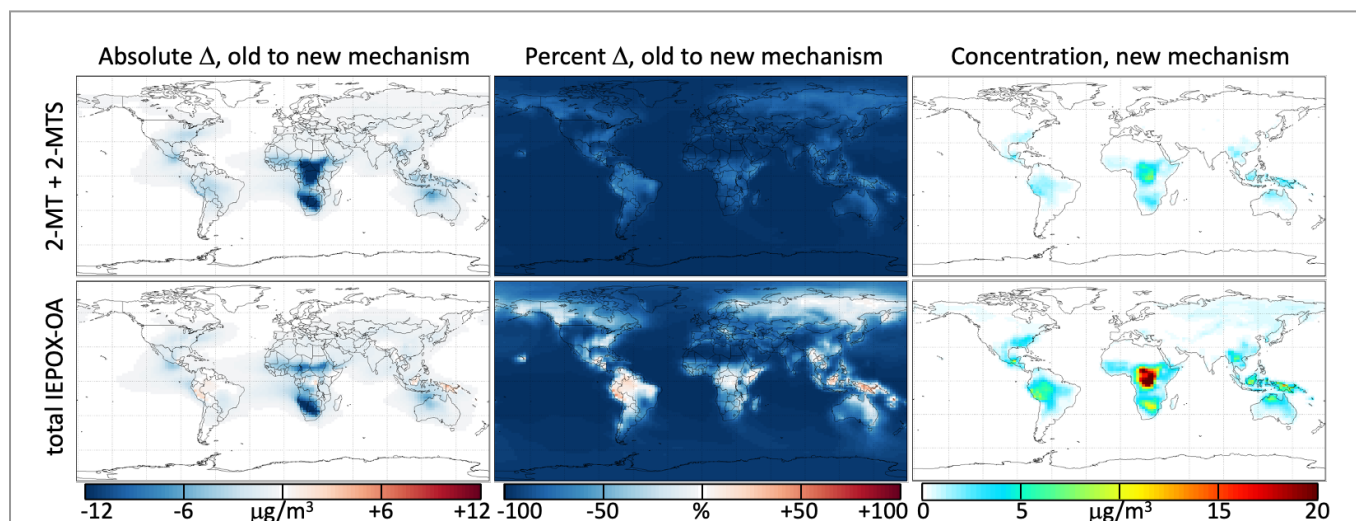
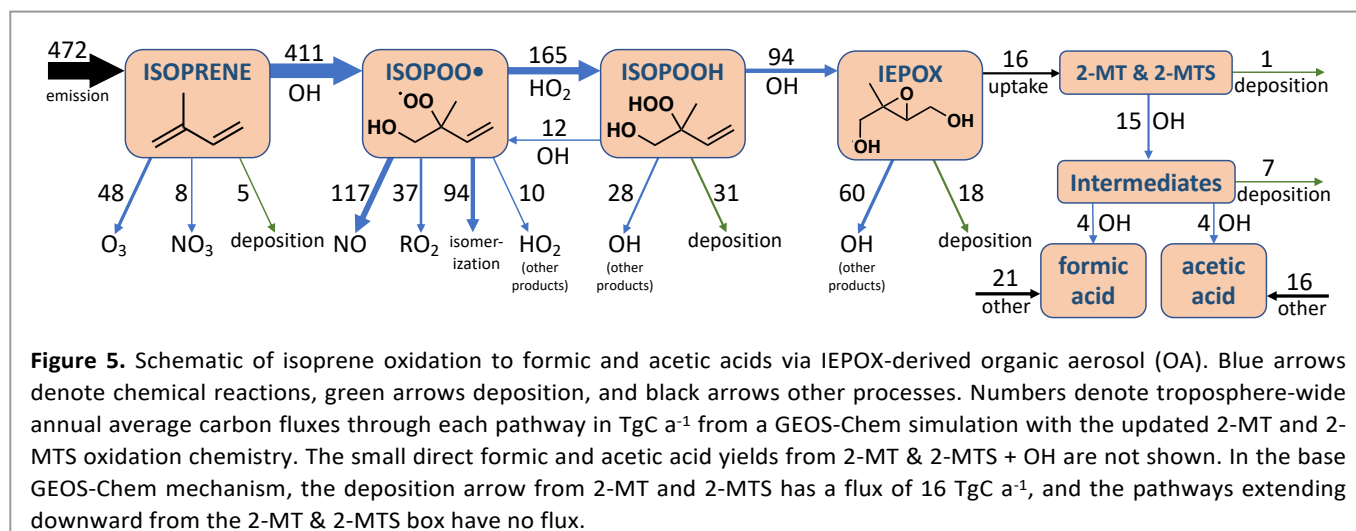


Figure 4. IEPOX-derived secondary organic aerosol (IEPOX-SOA) and changes in IEPOX-SOA concentrations between the base GEOS-Chem mechanism and the updated mechanism with 2-MT and 2-MTS oxidation. Panels show the absolute change between the mechanisms (left), the percent changes between the mechanisms (center), and the absolute concentrations in the updated mechanism (right) for both 2-MT + 2-MTS concentrations (top) and total IEPOX-SOA concentrations (bottom). All maps show annual averages at 0–1 km altitude.



model (Mechanism S1) used to simulate the photochemical experiments of both 2-MT (Cope *et al.*⁶³) and 2-MTS (this work) are incorporated into GEOS-Chem. The initial reactions of 2-MT and 2-MTS with OH are fast compared with deposition; as a result, the simulated total tropospheric mass burden of 2-MT + 2-MTS is reduced by 90% when their oxidation by OH is included in GEOS-Chem. Their aqueous-phase oxidation products, however, have longer lifetimes, so the total tropospheric burden of IEPOX-SOA only decreases by 67% with the new mechanism. Global surface values, which generally reflect more recently formed SOA with lower OH exposure, exhibit smaller declines – an 85% total reduction in 2-MT + 2-MTS and 47% total reduction in IEPOX-SOA. Results for the free troposphere (Figure S4) are spatially similar to those shown in Figure 4, but with much lower total burdens and slightly greater fractional decreases in 2-MT, 2-MTS, and IEPOX-SOA due to longer OH exposures. Seasonal variability in these changes, shown in Figures S5–S10, predominantly reflects the seasonal patterns of isoprene emissions.

Regional variability in the effects of the added 2-MT and 2-MTS chemistry reveals the extent of oxidation within particles, *i.e.*, the number of oxidative generations away from the C_5 parent compounds. The slightly more massive first-generation oxidative intermediates (see Figure 3 and Mechanism S1) can lead to a slight increase in total IEPOX-SOA mass relative to the base mechanism in locations close to IEPOX source regions. For example, in the Amazon, where OH is low due to titration by biogenic hydrocarbons, annual average surface IEPOX-SOA mass increases by 11% in simulations with the new 2-MT and 2-MTS chemistry. The second and third oxidation steps of 2-MT and 2-MTS lead to rapid organic aerosol mass loss and production of formic and acetic acids, which return to the gas phase. Therefore, in other isoprene-rich regions like the Congo basin and Southeast United States, where higher OH concentrations mean that more IEPOX-SOA oxidation is accomplished close to the source, annual average surface IEPOX-SOA mass decreases by 5% and 24% respectively in simulations with the new mechanism. Over the Southeast United States in the summer, this reduces the simulated boundary layer organic aerosol burden from 7.6 to $6.5 \mu\text{g m}^{-3}$,

improving the GEOS-Chem bias relative to atmospheric observations of OA by 60%.⁶⁸ Even sharper relative decreases in IEPOX-SOA burdens are evident in regions far from IEPOX-SOA sources, after multiple generations of photooxidation have occurred, where positive model biases have also been noted.⁶⁸

As shown in Figure 5, this added photochemical sink of 2-MT and 2-MTS in GEOS-Chem drastically changes the simulated IEPOX-SOA budget. Where the model previously predicted that all particle-phase 2-MT and 2-MTS (of which 16 Tg of carbon mass is produced annually) deposited unreacted, the updated model suggests that 92% of 2-MT and 2-MTS instead react with OH, forming more oxidized organic aerosol components and fragmenting to produce formic and acetic acids. These oxidized intermediates continue to react in the aqueous phase, and as a result, half of all carbon that enters the particle phase as IEPOX eventually becomes formic or acetic acids. The reactive flux through these oxidation pathways can also act as a major local sink of OH; the updated simulation shows decreased boundary layer OH of up to 15% compared to the base simulation over major isoprene-emitting regions (Figure S11).

Figure 6 shows the global distribution of formic acid produced by 2-MT and 2-MTS oxidation and the contribution of this source to the total atmospheric burden. As implemented in GEOS-Chem, the oxidation of IEPOX-SOA produces 14 Tg a^{-1} of formic acid – approximately equivalent to direct emissions from fires, anthropogenic sources, and plants combined (see Table S1) – increasing the global atmospheric source (emissions plus secondary production) by 18%. This in turn increases the tropospheric formic acid burden by 15%. These changes are spatially heterogeneous, with the largest absolute and fractional contributions of formic acid from IEPOX-SOA chemistry occurring over or downwind of isoprene source regions. Regionally, annual average surface formic acid from IEPOX-SOA chemistry can exceed 500 ppt, contributing over 50% to the local burden.

Because the majority of acid production takes place only after multiple aqueous oxidation steps, source areas with low OH such as the Amazon see minimal formic acid production from this mechanism; instead, most acid production occurs downwind. In contrast, the Southeast United States and Congo

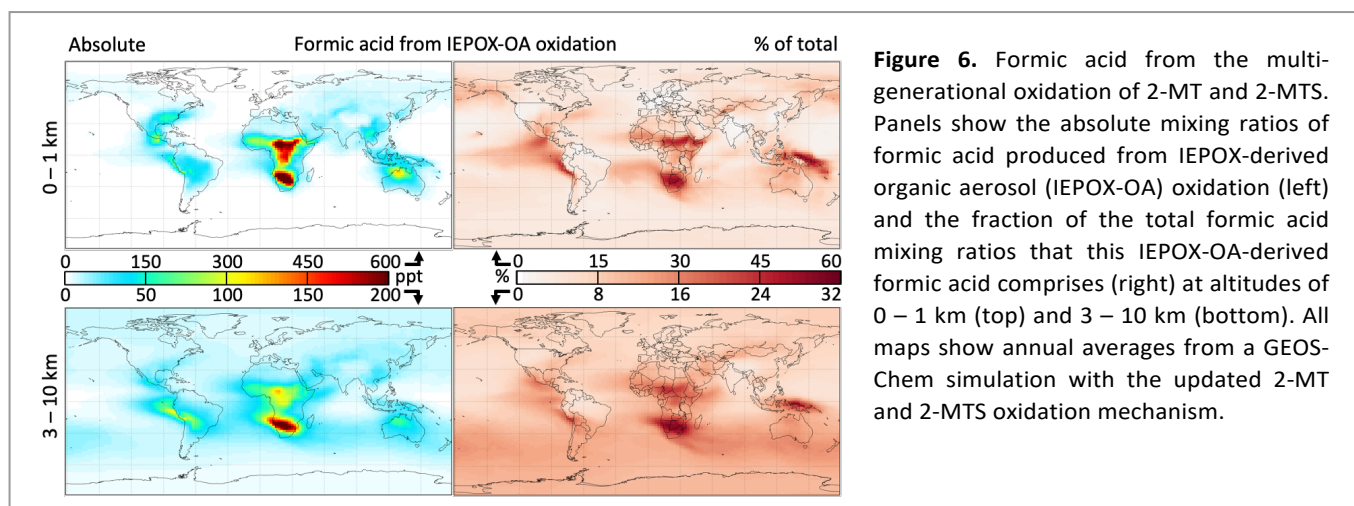


Figure 6. Formic acid from the multi-generational oxidation of 2-MT and 2-MTS. Panels show the absolute mixing ratios of formic acid produced from IEPOX-derived organic aerosol (IEPOX-OA) oxidation (left) and the fraction of the total formic acid mixing ratios that this IEPOX-OA-derived formic acid comprises (right) at altitudes of 0–1 km (top) and 3–10 km (bottom). All maps show annual averages from a GEOS-Chem simulation with the updated 2-MT and 2-MTS oxidation mechanism.

Basin sustain sufficiently high OH for 2-MT and 2-MTS to undergo multiple oxidation steps, resulting in high local formic acid production. Over the Southeast United States, where Millet *et al.*³⁹ showed that simulated near-surface formic acid is underestimated in GEOS-Chem by 1.5 ppb in the summer, the chemistry of 2-MT and 2-MTS produces an additional 500 pptv. IEPOX-SOA oxidation contributes to forested tropical Africa exhibiting some of the highest total formic acid mixing ratios in the world (Figure S12) sustained year-round (Figure S13), consistent with satellite observations.⁸⁷

Acetic acid production from IEPOX-SOA oxidation is approximately concurrent with that of formic acid; its distribution (Figures S14–16) therefore looks nearly identical to that shown in Figure 5. Although 2-MT and 2-MTS oxidation produces slightly less acetic acid (9.5 Tg a^{-1}) than formic acid, this makes up a larger fraction of the total atmospheric acetic acid source. The effect of this chemistry on the tropospheric acetic acid burden – a 19% increase – is therefore larger than the effect for formic acid, although the additional 300 pptv of acetic acid simulated from this chemistry over the Southeast United States in the summer is insufficient to make up the 1.1 ppb shortfall previously identified in GEOS-Chem.³⁹

The modeling results described above rely on a number of assumptions. First, we assume that the distribution of 2-MTS isomers present in the atmosphere reacts similarly to the one synthesized and studied here, which is likely only a minor isomer in the ambient distribution.⁸⁸ Second, our experiments were performed in dilute aqueous solution, whereas the atmospheric aerosols in which these reactions occur frequently have high ionic strength and likely contain many other possible coreactants. Third, we assume that 2-MT and 2-MTS and their IEPOX precursors do not undergo other reactions in aerosol, such as dimerization and oligomerization, that are not included in GEOS-Chem, although these processes have been shown to occur.^{51,88–90} While such oligomers could likely fragment upon OH-initiated oxidation as readily as their C_5 precursors, they may produce formic and acetic acids in different yields or timescales. Further mechanistic insights into the chemical pathways to formic and acetic acid formation in particle liquid water will be helpful for modeling.

A large source of uncertainty in these modeling results is the representation of particle-phase OH. GEOS-Chem simulates aqueous OH in aerosols as though it exists in rapid equilibrium with the gas phase, using a parameterization developed for cloud water.⁸⁰ Although this is certainly an oversimplification, aqueous atmospheric OH is notoriously difficult both to measure and to accurately model.^{91,92} Ambient particles are generally highly concentrated solutions containing many other species with which OH can react, in addition to numerous aqueous-phase sources of OH,^{93–96} none of which are explicitly represented here. While aqueous OH concentrations in GEOS-Chem ($\sim 10^{-13} \text{ M}$) are comparable to those in some other models,^{97–101} few measurements exist for comparison, and only for marine aerosol^{102–103} and larger droplets.⁹¹ Improving model representation of aerosol-phase OH will be critical for accurately simulating the fate of IEPOX-SOA in the atmosphere.

This study, with its noted assumptions, shows that adding photochemical sinks to SOA species in atmospheric models may have significant impacts on atmospheric composition. Although the cumulative effects of our assumptions on the atmospheric IEPOX-SOA oxidation rate and the magnitude of acid production cannot be straightforwardly determined, the results presented here provide a starting point from which to consider the potential impacts of aqueous-phase SOA photooxidation on the organic aerosol and organic acid budgets. In addition to this homogeneous (one phase) aging process, heterogeneous aging of IEPOX-SOA – the direct reaction of SOA with gas-phase OH – may also contribute. Studies of heterogeneous IEPOX-SOA aging using oxidation flow reactors (OFRs) measured much longer lifetimes (~ 2 weeks) of SOA against oxidative mass loss.^{104–106} However, such methods require their own assumptions to generalize from laboratory to ambient conditions; for example, the OH reactive-diffusive distance in particles is small,¹⁰⁷ and the high OH and short residence time in OFRs can make surface reactions self-limiting when the particle-phase organics cannot rapidly reach the surface,^{108–109} an effect exacerbated when the particles are highly organic.^{110–111} The actual atmospheric reactivity of IEPOX-SOA will depend on both gas- and particle-phase sources and sinks of OH, and on the aerosol phase, viscosity, water content, and chemical composition. Further

studies of aerosol oxidative aging in particle water at different particle phases and viscosities, and with aqueous OH sources and conditions relevant to the atmosphere, would facilitate an improved extrapolation from the lab to the ambient and more accurately inform future modeling studies.

Conclusions

We have shown using a synthetic standard in NMR-monitored reaction experiments that the aqueous-phase OH-initiated oxidation of 2-MTS, a ubiquitous product of IEPOX uptake into aerosol, rapidly produces formic and acetic acids in high yields in a multistep process. When incorporated into a global atmospheric model, simulations suggest that this oxidation (along with the analogous chemistry of 2-MT) is the dominant loss process of 2-MTS and 2-MT, reducing their combined tropospheric burden by 90%. These reactions may produce more oxidized aqueous intermediates that can locally increase SOA mass loadings, most notably over the Amazon, but ultimately lead to a 67% reduction in the tropospheric burden of total IEPOX-derived aerosol. The mass is removed from particles via conversion to gas-phase formic and acetic acids, increasing the tropospheric burdens of these species by 14.8% and 19.4% respectively. Aqueous-phase oxidation of IEPOX-SOA can also contribute significantly to local reactivity and depletion of OH, particularly over African forests.

While the modeling herein relies on many assumptions that introduce uncertainty to our results, we have shown that OH-initiated oxidation of aqueous IEPOX-SOA could substantially alter the global budgets of both organic aerosol and organic acids, and have provided a benchmark for quantifying these effects. Furthermore, the direction and magnitude of the changes to aerosol and acid burdens we simulate are in line with what previous studies suggest is needed to remedy model biases. Recent modeling studies suggest that incorporating all known sources of SOA from isoprene would lead to a severe overestimate of organic aerosol and that additional losses like those proposed here are needed to balance the SOA budget.^{36,57} Our simulated 1.1 $\mu\text{g m}^{-3}$ reduction in IEPOX-SOA over the Southeast United States in summer due to 2-MT and 2-MTS chemistry would reduce the current model-measurement bias in GEOS-Chem by 60%.⁶⁸ The increased tropospheric formic and acetic acid burdens from IEPOX-SOA oxidation also serve to reduce model-measurement biases; while the additional 500 ppt of formic acid and 300 ppt of acetic acid we simulate from 2-MT and 2-MTS chemistry over the Southeast United States in summer only makes up ~30% of the model shortfall,³⁹ similar oxidative chemistry of other particle-phase organics could further contribute. Extending the methods herein to other common aqueous aerosol constituents, improving model representation of aerosol-phase OH, and minimizing the other uncertainties from assumptions used in simulating gas-aerosol interactions will further improve our ability to simulate the burdens and budgets of organic aerosol and organic acids in the atmosphere.

Author contributions

KHB prepared and performed GEOS-Chem simulations with assistance from DJJ, XC, and DBM; JDC performed photooxidation experiments; KHB, JDC, and TBN analysed experimental data; KHB prepared the manuscript with assistance from all authors.

Conflicts of interest

The authors declare no conflicts of interest.

Acknowledgements

KHB and DJJ acknowledge support from the EPA STAR program (R840014). The views and opinions expressed in this article are those of the authors and do not represent the official views of the U.S. Environmental Protection Agency (EPA).

Notes and references

- 1 J. N. Galloway, G. E. Likens, W. C. Keene, and J. M. Miller, The composition of precipitation in remote areas of the world, *J. Geophys. Res.*, 1982, **87**, 8771–8786.
- 2 W. C. Keene and J. N. Galloway, Organic acidity in precipitation of North America, *Atmos. Environ.*, 1984, **18**, 2491–2497.
- 3 M. O. Andreae, R. W. Talbot, T. W. Andreae, and R. C. Harriss, Formic and acetic acid over the central Amazon region, Brazil: 1. Dry season, *J. Geophys. Res.*, 1988, **93**, 1616–1624.
- 4 P. Khare, N. Kumar, K. M. Kumari, and S. S. Srivastava, Atmospheric formic and acetic acids: An overview, *Rev. Geophys.*, 1999, **37** (2), 227–248.
- 5 F. Paulot, D. Wunch, J. D. Crouse, G. C. Toon, D. B. Millet, D. P. DeCarlo, C. Vigouroux, N. M. Deutscher, G. Gonzalez Abad, J. Notholt, T. Warneke, J. W. Hannigan, C. Warneke, J. A. de Gouw, E. J. Dunlea, M. De Maziere, D. W. T. Griffith, P. Bernath, J. L. Jimenez, and P. O. Wennberg, Importance of secondary sources in the atmospheric budgets of formic and acetic acids, *Atmos. Chem. Phys.*, 2011, **11** (5), 1989–2013.
- 6 V. Shah, D. J. Jacob, J. M. Moch, X. Wang, and S. Zhai, Global modeling of cloud water acidity, precipitation acidity, and acid inputs to ecosystems, *Atmos. Chem. Phys.*, 2020, **20**, 12223–12245.
- 7 S. C. Yu, Role of organic acids (formic, acetic, pyruvic and oxalic) in the formation of cloud condensation nuclei (CCN): a review, *Atmos. Res.*, 2000, **53** (4), 185–217.
- 8 J. D. Surratt, M. Lewandowski, J. H. Offenberg, M. Jaoui, T. E. Kleindienst, E. O. Edney, and J. H. Seinfeld, Effect of acidity on secondary organic aerosol formation from isoprene, *Environ. Sci. Technol.*, 2007, **41**, 5363–5369.
- 9 B. Ervens, B. J. Turpin, and R. J. Weber, Secondary organic aerosol formation in cloud droplets and aqueous particles (aqSOA): a review of laboratory, field and model studies, *Atmos. Chem. Phys.*, 2011, **11**, 11069–11102.
- 10 J. Kesselmeier, K. Bode, C. Gerlach, and E. M. Jork, Exchange of atmospheric formic and acetic acids with trees and crop plants under controlled chamber and purified air conditions, *Atmos. Environ.*, 1998, **32** (10), 1765–1775.
- 11 A. Guenther, C. Geron, T. Pierce, B. Lamb, P. Harley, and R. Fall, Natural emissions of non-methane volatile organic compounds, carbon monoxide, and oxides of nitrogen from North America, *Atmos. Environ.*, 2000, **34**, 2205–2230.
- 12 U. Kuhn, S. Rottenberger, T. Biesenthal, C. Ammann, A. Wolf, G. Schebeske, S. T. Oliva, T. M. Tavares, and J. Kesselmeier,

- Exchange of short-chain monocarboxylic acids by vegetation at a remote tropical forest site in Amazonia, *J. Geophys. Res. - Atmos.*, 2002, **107**, 8069.
- 13 A. Ito and J. E. Penner, Global estimates of biomass burning emissions based on satellite imagery for the year 2000, *J. Geophys. Res. - Atmos.*, 2004, **109**, D14S05.
- 14 P. Veres, J. M. Roberts, I. R. Burling, C. Warneke, J. de Gouw, and R. J. Yokelson, Measurements of gas-phase inorganic and organic acids from biomass fires by negative-ion proton-transfer chemical-ionization mass spectrometry, *J. Geophys. Res. - Atmos.*, 2010, **115**, D23302.
- 15 S. Chaliyakunnel, D. B. Millet, K. C. Wells, K. E. Cady-Pereira, and M. W. Shephard, A large underestimate of formic acid from tropical fires: Constraints from space-borne measurements, *Environ. Sci. Technol.*, 2016, **50**, 5631-5640.
- 16 E. Sanhueza and M. O. Andreae, Emission of formic and acetic acids from tropical Savanna soils, *Geophys. Res. Lett.*, 1991, **18**, 1707-1710.
- 17 B. T. Jones, J. Muller, S. O'Shea, A. Bacak, G. Allen, M. Gallagher, K. Bower, M. Le Breton, T. J. Bannan, S. Bauguitte, J. Pyle, D. Lowry, R. Fisher, J. France, E. Nisbet, D. Shallcross, and C. Percival, Are the Fenno-Scandinavian arctic wetlands a significant regional source of formic acid?, *Atmosphere*, 2017, **8**, 112.
- 18 A. Mielnik, M. Link, J. Mattila, S. R. Fulgham, and D. K. Farmer, Emission of formic and acetic acid from two Colorado soils. *Environ. Sci.: Processes & Impacts*, 2018, **20**, 1537-1545.
- 19 J. E. Dibb and M. Arsenault, Shouldn't snowpacks be sources of monocarboxylic acids?, *Atmos. Environ.*, 2002, **36**, 2513-2522.
- 20 K. Kawamura, L. L. Ng, and I. R. Kaplan, Determination of organic acids (C1-C10) in the atmosphere, motor exhausts, and engine oils, *Environ. Sci. Technol.*, 1985, **19**, 1082-1086.
- 21 T. J. Bannan, A. Bacak, J. B. A. Muller, A. M. Booth, B. Jones, M. Le Breton, K. E. Leather, M. Ghalaieny, P. Xiao, D. E. Shallcross, and C. J. Percival, Importance of direct anthropogenic emissions of formic acid measured by a chemical ionisation mass spectrometer (CIMS) during the Winter ClearLo Campaign in London, January 2012, *Atmos. Environ.*, 2014, **83**, 301-310.
- 22 T. A. Crisp, J. M. Brady, C. D. Cappa, S. Collier, S. D. Forestieri, M. J. Kleeman, T. Kuwayama, B. M. Lerner, E. J. Williams, Q. Zhang, and T. H. Bertram, On the primary emission of formic acid from light duty gasoline vehicles and ocean-going vessels, *Atmos. Environ.*, 2014, **98**, 426-433.
- 23 J. Liggio, S. G. Moussa, J. Wentzell, A. Darlington, P. Liu, A. Leithead, K. Hayden, J. O'Brien, R. L. Mittermeier, R. Staebler, M. Wolde, and S.-M. Li, Understanding the primary emissions and secondary formation of gaseous organic acids in the oil sands region of Alberta, Canada, *Atmos. Chem. Phys.*, 2017, **17**, 8411-8427.
- 24 B. R. Larsen, D. Di Bella, M. Glasius, R. Winterhalter, N. Jensen, and J. Hjorth, Gas-phase OH oxidation of monoterpenes: Gaseous and particulate products. *J. Atmos. Chem.*, 2001, **38** (3), 231-276.
- 25 A. Lee, A. H. Goldstein, J. H. Kroll, N. L. Ng, V. Varutbangkul, R. C. Flagan, and J. H. Seinfeld, Gas-phase products and secondary aerosol yields from the photooxidation of 16 different terpenes. *J. Geophys. Res.*, 2006, **111**, D17305.
- 26 T. B. Nguyen, G. S. Tyndall, J. D. Crouse, A. P. Teng, K. H. Bates, R. H. Schwantes, M. M. Coggon, L. Zhang, P. Feiner, D. O. Miller, K. M. Skog, J. C. Rivera-Rios, M. Dorris, K. F. Olson, A. Koss, R. J. Wild, S. S. Brown, A. H. Goldstein, J. A. de Gouw, W. H. Brune, F. N. Keutsch, J. H. Seinfeld, and P. O. Wennberg, Atmospheric fate of Criegee intermediates in the ozonolysis of isoprene. *Phys. Chem. Chem. Phys.*, 2016, **18** (15), 10241-10254.
- 27 H. M. Allen, J. D. Crouse, K. H. Bates, A. P. Teng, M. P. Krawiec-Thayer, J. C. Rivera-Rios, F. N. Keutsch, J. M. St. Clair, T. F. Hanisco, K. H. Møller, H. G. Kjaergaard, and P. O. Wennberg, Kinetics and product yields of the OH initiated oxidation of hydroxymethyl hydroperoxide. *J. Phys. Chem. A*, 2018, **122**, 6292-6302.
- 28 H. D. Alwe, D. B. Millet, X. Chen, J. D. Raff, Z. C. Payne, and K. Fledderman, Oxidation of volatile organic compounds as the major source of formic acid in a mixed forest canopy, *Geophys. Res. Lett.*, 2019, **46**, 2940-2948.
- 29 M. F. Link, T. B. Nguyen, K. H. Bates, J.-F. Müller, and D. K. Farmer, Can isoprene oxidation explain high concentrations of atmospheric formic and acetic acid over forests? *ACS Earth Space Chem.*, 2020, **4** (5), 730-740.
- 30 J. G. Goode, R. J. Yokelson, D. E. Ward, R. A. Susott, R. E. Babbitt, M. A. Davies, and W. M. Hao, Measurements of excess O₃, CO₂, CH₄, C₂H₄, C₂H₂, HCN, NO, NH₃, HCOOH, CH₃COOH, HCHO, and CH₃OH in 1997 Alaskan biomass burning plumes by airborne Fourier transform infrared spectroscopy (AFTIR), *J. Geophys. Res.*, 2000, **105**, 22147-22166.
- 31 S. Gao, D. Hegg, P. Hobbs, T. Kirchstetter, B. Magi, and M. Sadilek, Water-soluble organic components in aerosols associated with savanna fires in southern Africa: Identification, evolution, and distribution, *J. Geophys. Res.*, 2003, **108**, 8491.
- 32 R. J. Yokelson, I. T. Bertschi, T. J. Christian, P. V. Hobbs, D. E. Ward, and W. M. Hao: Trace gas measurements in nascent, aged, and cloud-processed smoke from African savanna fires by airborne Fourier transform infrared spectroscopy (AFTIR), *J. Geophys. Res.*, 2003, **108**, 8478.
- 33 S. C. Herndon, M. S. Zahniser, D. D. Nelson, J. Shorter, J. B. McManus, R. Jiménez, C. Warneke, and J. A. de Gouw, Airborne measurements of HCHO and HCOOH during the New England Air Quality Study 2004 using a pulsed quantum cascade laser spectrometer, *J. Geophys. Res. - Atmos.*, 2007, **112**, D10S03.
- 34 B. Yuan, P. R. Veres, C. Warneke, J. M. Roberts, J. B. Gilman, A. Koss, P. M. Edwards, M. Graus, W. C. Kuster, S.-M. Li, R. J. Wild, S. S. Brown, W. P. Dubé, B. M. Lerner, E. J. Williams, J. E. Johnson, P. K. Quinn, T. S. Bates, B. Lefer, P. L. Hayes, J. L. Jimenez, R. J. Weber, R. Zamora, B. Ervens, D. B. Millet, B. Rappenglück, and J. A. de Gouw, Investigation of secondary formation of formic acid: urban environment vs. oil and gas producing region, *Atmos. Chem. Phys.*, 2015, **15**, 1975-1993.
- 35 S. Wang, M. J. Newland, W. Deng, A. R. Rickard, J. F. Hamilton, A. Munoz, M. Rodenas, M. M. Vazquez, L. Wang, and X. Wang, Aromatic photo-oxidation, a new source of atmospheric acidity. *Environ. Sci. Technol.*, 2020, **54** (13), 7798-7806.
- 36 K. H. Bates, D. J. Jacob, K. Li, P. D. Ivatt, M. J. Evans, Y. Yan, and J. Lin, Development and evaluation of a new compact mechanism for aromatic oxidation in atmospheric models. *Atmos. Chem. Phys.*, 2021, **21**, 18351-18374.
- 37 T. Stavrou, J.-F. Müller, J. Peeters, A. Razavi, L. Clarisse, C. Clerbaux, P.-F. Coheur, D. Hurtmans, M. D. Mazière, C. Vigouroux, N. M. Deutscher, D. W. T. Griffith, N. Jones, and C. Paton-Walsh, Satellite evidence for a large source of formic acid from boreal and tropical forests, *Nat. Geosci.*, 2012, **5**, 26-30.
- 38 K. E. Cady-Pereira, S. Chaliyakunnel, M. W. Shephard, D. B. Millet, M. Luo, and K. C. Wells, HCOOH measurements from space: TES retrieval algorithm and observed global distribution, *Atmos. Meas. Tech.*, 2014, **7**, 2297-2311.
- 39 D. B. Millet, M. Baasandorj, D. K. Farmer, J. A. Thornton, K. Baumann, P. Brophy, S. Chaliyakunnel, J. A. de Gouw, M. Graus, L. Hu, A. Koss, B. H. Lee, F. D. Lopez-Hilfiker, J. A. Neuman, F. Paulot, J. Peischl, I. B. Pollack, T. B. Ryerson, C. Warneke, B. J. Williams, and J. Xu, A large and ubiquitous

- source of atmospheric formic acid. *Atmos. Chem. Phys.* 2015, **15** (11), 6283–6304.
- 40 X. Chen, D. B. Millet, H. B. Singh, A. Wisthaler, E. C. Apel, E. L. Atlas, D. R. Blake, I. Bourgeois, S. S. Brown, J. D. Crouse, J. A. de Gouw, F. M. Flocke, A. Fried, B. G. Heikes, R. S. Hornbrook, T. Mikoviny, K. E. Min, M. Müller, J. A. Neuman, D. W. O'Sullivan, J. Peischl, G. G. Pfister, D. Richter, J. M. Roberts, T. B. Ryerson, S. R. Shertz, C. R. Thompson, V. Treadaway, P. R. Veres, J. Walega, C. Warneke, R. A. Washenfelder, P. Weibring, and B. Yuan, On the sources and sinks of atmospheric VOCs: an integrated analysis of recent aircraft campaigns over North America. *Atmos. Chem. Phys.* 2019, **19** (14), 9097–9123.
- 41 D. J. Jacob, Chemistry of OH in remote clouds and its role in the production of formic acid and peroxymonosulfate, *J. Geophys. Res.*, 1986, **91**, 9807–9826.
- 42 T. L. Eliason, S. Aloisio, D. J. Donaldson, D. J. Cziczko, and V. Vaida, Processing of unsaturated organic acid films and aerosols by ozone, *Atmos. Environ.*, 2003, **37**, 2207–2219.
- 43 M. J. Molina, A. V. Ivanov, S. Trakhtenberg, and L. T. Molina, Atmospheric evolution of organic aerosol, *Geophys. Res. Lett.*, 2004, **31**, L22104.
- 44 M. L. Walsler, J. Park, A. L. Gomez, A. R. Russell, and S. A. Nizkorodov, Photochemical aging of secondary organic aerosol particles generated from the oxidation of d-limonene, *J. Phys. Chem. A*, 2007, **111**, 1907–1913.
- 45 A. Vlasenko, I. J. George, and J. P. D. Abbatt, Formation of volatile organic compounds in the heterogeneous oxidation of condensed-phase organic films by gas-phase OH, *J. Phys. Chem. A*, 2008, **112**, 1552–1560.
- 46 X. Pan, J. S. Underwood, J.-H. Xing, S. A. Mang, and S. A. Nizkorodov, Photodegradation of secondary organic aerosol generated from limonene oxidation by ozone studied with chemical ionization mass spectrometry, *Atmos. Chem. Phys.*, 2009, **9**, 3851–3865.
- 47 B. Franco, T. Blumenstock, C. Cho, C. Clarisse, C. Cierbaux, P.-F. Coheur, M. de Maziere, I. de Smedt, H.-P. Dorn, T. Emmerichs, H. Fuchs, G. Gkatzelis, D. W. T. Griffith, S. Gromov, J. W. Hannigan, F. Hase, T. Hohaus, N. Jones, A. Kerkweg, A. Kiendler-Scharr, E. Lutsch, E. Mahieu, A. Novelli, I. Ortega, C. Paton-Walsh, M. Pommier, A. Pozzer, D. Reimer, S. Rosanka, R. Sander, M. Schneider, K. Strong, R. Tillmann, M. van Roozendaal, L. Vereecken, C. Vigouroux, A. Wahner, and D. Taraborrelli, Ubiquitous atmospheric production of organic acids mediated by cloud droplets. *Nature*, 2021, **593**, 233–237.
- 48 J. D. Surratt, A. W. H. Chan, N. C. Eddingsaas, M. N. Chan, C. L. Loza, A. J. Kwan, S. P. Hersey, R. C. Flagan, P. O. Wennberg, and J. H. Seinfeld, Reactive intermediates revealed in secondary organic aerosol formation from isoprene, *Proc. Nat'l. Acad. Sci.*, 2010, **107**, 6640–6645.
- 49 T. B. Nguyen, M. M. Coggon, K. H. Bates, X. Zhang, R. H. Schwantes, K. A. Schilling, C. L. Loza, R. C. Flagan, P. O. Wennberg, and J. H. Seinfeld, Organic aerosol formation from the reactive uptake of isoprene epoxydiols (IEPOX) onto non-acidified inorganic seeds, *Atmos. Chem. Phys.*, 2014, **14**, 3497–3510.
- 50 J. D. Surratt, J. H. Kroll, T. E. Kleindienst, E. O. Edney, M. Claeys, A. Sorooshian, N. L. Ng, J. H. Offenberg, M. Lewandowski, M. Jaoui, R. C. Flagan, and J. H. Seinfeld, Evidence for organosulfates in secondary organic aerosol, *Environ. Sci. Technol.*, 2007, **41**, 517–527.
- 51 Y. H. Lin, Z. F. Zhang, K. S. Docherty, H. F. Zhang, S. H. Budisulistiorini, C. L. Rubitschun, S. L. Shaw, E. M. Knipping, E. S. Edgerton, T. E. Kleindienst, A. Gold, and J. D. Surratt, Isoprene epoxydiols as precursors to secondary organic aerosol formation: Acid-catalyzed reactive uptake studies with authentic compounds, *Environ. Sci. Technol.*, 2012, **46**, 250–258.
- 52 I. Kourttchev, T. Ruuskanen, W. Maenhaut, M. Kulmala, and M. Claeys, Observation of 2-methyltetrols and related photo-oxidation products of isoprene in boreal forest aerosols from Hyytiälä, Finland. *Atmos. Chem. Phys.*, 2005, **5** (10), 2761–2770.
- 53 A. L. Clements and J. H. Seinfeld, Detection and quantification of 2-methyltetrols in ambient aerosol in the southeastern United States. *Atmos. Environ.*, 2007, **41** (9), 1825–1830.
- 54 L. L. Liang, G. Engling, F. K. Duan, Y. Cheng, and K. B. He, Characteristics of 2-methyltetrols in ambient aerosol in Beijing, China. *Atmos. Environ.*, 2012, **59**, 376–381.
- 55 J. Liao, K. D. Froyd, D. M. Murphy, F. N. Keutsch, G. Yu, P. O. Wennberg, J. M. St. Clair, J. D. Crouse, A. Wisthaler, T. Mikoviny, J. L. Jimenez, P. Campuzano-Jost, D. A. Day, W. Hu, T. B. Ryerson, I. B. Pollack, J. Peischl, B. E. Anderson, L. D. Ziemba, D. R. Blake, S. Meinardi, and G. Diskin, Airborne measurements of organosulfates over the continental US, *J. Geophys. Res. - Atmos.*, 2015, **120** (7), 2990–3005
- 56 D. D. Hughes, M. B. Christiansen, A. Milani, M. P. Vermeuel, G. A. Novak, H. D. Alwe, A. F. Dickens, R. B. Pierce, D. B. Millet, T. H. Bertram, C. O. Stanier, and E. A. Stone, PM2.5 chemistry, organosulfates, and secondary organic aerosol during the 2017 Lake Michigan Ozone Study. *Atmos. Environ.*, 2021, **244**, 117939.
- 57 S. Stadler, T. Kühn, S. Schröder, D. Taraborrelli, M. G. Schultz, and H. Kokkola, Isoprene-derived secondary organic aerosol in the global aerosol–chemistry–climate model ECHAM6.3.0–HAM2.3–MOZ1.0, *Geosci. Model Dev.*, 2018, **11**, 3235–3260.
- 58 K. H. Bates and D. J. Jacob, A new model mechanism for atmospheric oxidation of isoprene: global effects on oxidants, nitrogen oxides, organic products, and secondary organic aerosol, *Atmos. Chem. Phys.*, 2019, **19**, 9613–9640.
- 59 A. Hodzic, P. S. Kasibhatla, D. S. Jo, C. D. Cappa, J. L. Jimenez, S. Madronich, and R. J. Park, Rethinking the global secondary organic aerosol (SOA) budget: stronger production, faster removal, shorter lifetime, *Atmos. Chem. Phys.*, 2016, **16**, 7917–7941.
- 60 A. Hodzic, S. Madronich, P. S. Kasibhatla, G. Tyndall, B. Aumont, J. L. Jimenez, J. Lee-Taylor, and J. Orlando, Organic photolysis reactions in tropospheric aerosols: effect on secondary organic aerosol formation and lifetime, *Atmos. Chem. Phys.*, 2015, **15**, 9253–9269.
- 61 M. A. Zawadowicz, B. H. Lee, M. Shrivastava, A. Zelenyuk, R. A. Zaveri, C. Flynn, J. A. Thornton, and J. E. Shilling, Photolysis controls atmospheric budgets of biogenic secondary organic aerosol. *Environ. Sci. Technol.*, 2020, **54** (7), 3861–3870.
- 62 S. Lou, M. Shrivastava, R. C. Easter, Y. Yang, P.-L. Ma, H. Wang, M. J. Cubison, P. Campuzano-Jost, J. L. Jimenez, Q. Zhang, P. J. Rasch, J. E. Shilling, A. Zelenyuk, M. Dubey, P. Cameron-Smith, S. T. Martin, J. Schneider, and C. Schulz, New SOA treatments within the Energy Exascale Earth System Model (E3SM): Strong production and sinks govern atmospheric SOA distributions and radiative forcing. *J. Adv. Model. Earth Syst.*, 2020, **12**, e2020MS002266.
- 63 J. D. Cope, K. A. Abellar, K. H. Bates, X. Fu, and T. B. Nguyen, Aqueous photochemistry of 2-methyltetrol and erythritol as sources of formic acid and acetic acid in the atmosphere. *ACS Earth Space Chem.*, 2021, **5** (6), 1265–1277.
- 64 A. L. Bondy, R. L. Craig, Z. Zhang, A. Gold, J. D. Surratt, and A. P. Ault, Isoprene-Derived Organosulfates: Vibrational Mode Analysis by Raman Spectroscopy, Acidity-Dependent Spectral Modes, and Observation in Individual Atmospheric Particles. *J. Phys. Chem. A*, 2018, **122** (1), 303–315.
- 65 K. A. Abellar, J. D. Cope, and T. B. Nguyen, Second-order kinetic rate coefficients for the aqueous-phase hydroxyl radical (OH) oxidation of isoprene-derived secondary organic

- aerosol compounds at 298 K. *Environ. Sci. Technol.*, 2021, **55** (20), 13728–13736.
- 66 J. B. Burkholder, S. P. Sander, J. Abbatt, J. R. Barker, C. Cappa, J. D. Crouse, T. S. Dibble, R. E. Huie, C. E. Kolb, M. J. Kurylo, V. L. Orkin, C. J. Percival, D. M. Willmouth, and P. H. Wine, Chemical Kinetics and Photochemical Data for Use in Atmospheric Studies, Evaluation No. 19, JPL Publication 19-5, Jet Propulsion Laboratory, Pasadena, 2019, <http://jpldataeval.jpl.nasa.gov>.
- 67 M. Chin and P. H. Wine, A temperature-dependent competitive kinetics study of the aqueous-phase reactions of OH Radicals with Formate, Formic-Acid, Acetate, Acetic-Acid, and Hydrated Formaldehyde. *Aquatic and Surface Photochemistry*, 1994, 85–96.
- 68 S. J. Pai, C. L. Heald, J. R. Pierce, S. C. Farina, E. A. Marais, J. L. Jimenez, P. Campuzano-Jost, B. A. Nault, A. M. Middlebrook, H. Coe, J. E. Shilling, R. Bahreini, J. H. Dingle, and K. Vu, An evaluation of global organic aerosol schemes using airborne observations. *Atmos. Chem. Phys.*, 2020, **20**, 2637–2665.
- 69 S.-J. Lin and R. B. Rood, Multidimensional flux-form semi-Lagrangian transport schemes. *Mon. Weather Rev.*, 1996, **124**, 2046–2070.
- 70 J.-T. Lin and M. B. McElroy, Impacts of boundary layer mixing on pollutant vertical profiles in the lower troposphere: Implications to satellite remote sensing. *Atmos. Environ.*, 2010, **44**, 1726–1739.
- 71 C. A. Keller, M. S. Long, R. M. Yantosca, A. M. Da Silva, S. Pawson, and D. J. Jacob, HEMCO v1.0: a versatile, ESMF-compliant component for calculating emissions in atmospheric models. *Geosci. Model Dev.*, 2014, **7**, 1409–1417.
- 72 A. Guenther, X. Jiang, C. L. Heald, T. Sakulyanontvittaya, T. Duhl, L. K. Emmons, and X. Wang, The Model of Emissions of Gases and Aerosols from Nature version 2.1 (MEGAN 2.1): an extended and updated framework for modeling biogenic emissions. *Geosci. Model Dev.*, 2012, **5**, 1471–1492.
- 73 R. M. Hoesly, S. J. Smith, L. Feng, Z. Klimont, G. Janssens-Maenhout, T. Pitkanen, J. J. Seibert, L. Vu, R. J. Andres, R. M. Bolt, T. C. Bond, L. Dawidowski, N. Kholod, J.-I. Kurokawa, M. Li, L. Liu, Z. Lu, M. C. P. Moura, P. R. O'Rourke, and Q. Zhang, Historical (1750–2014) anthropogenic emissions of reactive gases and aerosols from the Community Emissions Data System (CEDS). *Geosci. Model Dev.*, 2018, **11**, 369–408.
- 74 G. R. van der Werf, J. T. Randerson, L. Giglio, G. J. Collatz, M. Mu, P. S. Kasibhatla, D. C. Morton, R. S. DeFries, Y. Jin, and T. T. van Leeuwen, Global fire emissions and the contribution of deforestation, savanna, forest, agricultural, and peat fires (1997–2009). *Atmos. Chem. Phys.*, 2010, **10**, 11707–11735.
- 75 E. A. Marais and C. Wiedinmyer, Air Quality Impact of Diffuse and Inefficient Combustion Emissions in Africa (DICE-Africa). *Environ. Sci. Technol.*, 2016, **50**, 10739–10745.
- 76 P. O. Wennberg, K. H. Bates, J. D. Crouse, L. G. Dodson, R. C. McVay, L. A. Mertens, T. B. Nguyen, E. Praske, R. H. Schwantes, M. D. Smarte, J. M. St. Clair, A. P. Teng, X. Zhang, and J. H. Seinfeld, The gas-phase oxidation of isoprene and its first-generation products. *Chem. Rev.*, 2018, **118** (7), 3337–3390.
- 77 E. A. Marais, D. J. Jacob, J. L. Jimenez, P. Campuzano-Jost, D. A. Day, W. Hu, J. Krechmer, L. Zhu, P. S. Kim, C. C. Miller, J. A. Fisher, K. Travis, K. Yu, T. F. Hanisco, G. M. Wolfe, H. L. Arkinson, H. O. T. Pye, K. D. Froyd, J. Liao, and V. F. McNeill, Aqueous-phase mechanism for secondary organic aerosol formation from isoprene: application to the southeast United States and co-benefit of SO₂ emission controls. *Atmos. Chem. Phys.*, 2016, **16**, 1603–1618.
- 78 N. C. Cole-Filipiak, A. E. O'Connor, and M. J. Elrod, Kinetics of the hydrolysis of atmospherically relevant isoprene-derived hydroxy epoxides. *Environ. Sci. Technol.*, 2010, **44**, 6718–6723.
- 79 N. C. Eddingsaas, D. G. VanderVelde, and P. O. Wennberg, Kinetics and products of the acid-catalyzed ring-opening of atmospherically relevant butyl epoxy alcohols. *J. Phys. Chem. A*, 2010, **114**, 8106–8113.
- 80 D. J. Jacob, B. D. Field, Q. Li, D. R. Blake, J. de Gouw, C. Warneke, A. Hansel, A. Wisthaler, H. B. Singh, and A. Guenther, Global budget of methanol: Constraints from atmospheric observations. *J. Geophys. Res.*, 2005, **110**, D08303.
- 81 X. Chen, D. B. Millet, J. A. Neuman, P. R. Veres, E. A. Ray, R. Commane, B. C. Daube, K. McKain, J. P. Schwarz, J. M. Katich, K. D. Froyd, G. P. Schill, M. J. Kim, J. D. Crouse, H. M. Allen, E. C. Apel, R. S. Hornbrook, D. R. Blake, B. A. Nault, P. Campuzano-Jost, J. L. Jimenez, and J. E. Dibb, HCOOH in the Remote Atmosphere: Constraints from Atmospheric Tomography (ATom) Airborne Observations. *ACS Earth Space Chem.*, 2021, **5** (6), 1436–1454.
- 82 M. F. Shaw, B. Sztáray, L. K. Whalley, D. E. Heard, D. B. Millet, M. J. T. Jordan, D. L. Osborn, and S. H. Kable, Phototautomerization of acetaldehyde as a photochemical source of formic acid in the troposphere. *Nat. Commun.*, 2018, **9**, 2584.
- 83 K. H. Bates, J. D. Cope, and T. B. Nguyen, Gas-phase oxidation rates and products 1,2-dihydroxy isoprene. *Environ. Sci. Technol.*, 2021, **55**, 14294–14304.
- 84 K. T. Vasquez, J. D. Crouse, B. C. Schulze, K. H. Bates, A. P. Teng, L. Xu, and P. O. Wennberg, Rapid hydrolysis of tertiary isoprene nitrate efficiently removes NO_x from the atmosphere. *Proc. Nat'l Acad. Sci.*, 2020, **117** (52) 33011–33016.
- 85 E. Dovrou, J. C. Rivera-Rios, K. H. Bates, and F. N. Keutsch, Sulfate formation via cloud processing from isoprene hydroxyl hydroperoxides (ISOPOOH). *Environ. Sci. Technol.*, 2019, **53** (21), 12476–12484.
- 86 E. Dovrou, K. H. Bates, J. C. Rivera-Rios, J. L. Cox, J. D. Shutter, and F. N. Keutsch, Towards a chemical mechanism of the oxidation of aqueous sulfur dioxide via isoprene hydroxyl hydroperoxides (ISOPOOH). *Atmos. Chem. Phys.*, 2021, **21**, 8999–9008.
- 87 A. Razavi, F. Karagulian, L. Clarisse, D. Hurtmans, P. F. Coheur, C. Clerbaux, J.-F. Müller, and T. Stavrou, Global distributions of methanol and formic acid retrieved for the first time from the IASI/MetOp thermal infrared sounder. *Atmos. Chem. Phys.*, 2011, **11**, 857–872.
- 88 T. Cui, Z. Zeng, E. O. dos Santos, Z. Zhang, Y. Chen, Y. Zhang, C. A. Rose, S. H. Budisulistiorini, L. B. Collins, W. M. Bodnar, R. A. F. de Souza, S. T. Martin, C. M. D. Machado, B. J. Turpin, A. Gold, A. P. Ault, and J. D. Surratt, Development of a hydrophilic interaction liquid chromatography (HILIC) method for the chemical characterization of water-soluble isoprene epoxydiol (IEPOX)-derived secondary organic aerosol. *Environ. Sci.-Proc. Imp.*, 2018, **20**, 1524–1536.
- 89 Y.-H. Lin, S. Budisulistiorini, K. Chu, R. A. Siejack, H. Zhang, M. Riva, Z. Zhang, A. Gold, K. E. Kautzman, and J. D. Surratt, Light-Absorbing Oligomer Formation in Secondary Organic Aerosol from Reactive Uptake of Isoprene Epoxydiols. *Environ. Sci. Technol.*, 2014, **48**, 12012–12021.
- 90 S. J. Stropoli, C. R. Miner, D. R. Hill, and M. J. Elrod, Assessing Potential Oligomerization Reaction Mechanisms of Isoprene Epoxydiols on Secondary Organic Aerosol. *Environ. Sci. Technol.*, 2018, **53**, 176–184.
- 91 Y. Rudich, N. M. Donahue, and T. F. Mentel, Aging of Organic Aerosol: Bridging the Gap Between Laboratory and Field Studies. *Ann. Rev. Phys. Chem.*, 2007, **58**, 321–352.
- 92 A. Tilgner and H. Herrmann, Tropospheric Aqueous-Phase OH Oxidation Chemistry: Current Understanding, Uptake of Highly Oxidized Organics and Its Effect, in *Multiphase Environmental Chemistry in the Atmosphere*, ed. S. W. Hunt, A. Laskin, and S. A. Nizkorodov, 2018, 49–85.

- 93 K. M. Badali, S. Zhou, D. Aljawhary, M. Antiñolo, W. J. Chen, A. Lok, E. Mungall, J. P. S. Wong, R. Zhao, and J. P. D. Abbatt, Formation of hydroxyl radicals from photolysis of secondary organic aerosol material, *Atmos. Chem. Phys.*, 2015, **15**, 7831–7840.
- 94 H. Tong, A. M. Arangio, P. S. J. Lakey, T. Berkemeier, F. Liu, C. J. Kampf, W. H. Brune, U. Pöschl, and M. Shiraiwa, Hydroxyl radicals from secondary organic aerosol decomposition in water, *Atmos. Chem. Phys.*, 2016, **16**, 1761–1771.
- 95 M. Gen, R. Zhang, D. D. Huang, Y. Li, and C. K. Chan, Heterogeneous SO₂ oxidation in sulfate formation by photolysis of particulate nitrate, *Environ. Sci. Technol. Lett.*, 2019, **6** (2), 86–91.
- 96 T. Arakaki, Y. Kuroki, K. Okada, Y. Nakama, H. Ikota, M. Kinjo, T. Higuchi, M. Uehara, and A. Tanahara, Chemical composition and photochemical formation of hydroxyl radicals in aqueous extracts of aerosol particles collected in Okinawa, Japan, *Atmos. Environ.*, 2006, **40** (25), 4764–4774.
- 97 H. Herrmann, D. Hoffmann, T. Schaefer, P. Bräuer, and A. Tilgner, Tropospheric aqueous-phase free-radical chemistry: radical sources, spectra, reaction kinetics and prediction tools, *ChemPhysChem.*, 2017, **11**, 3796–3822.
- 98 B. Ervens and R. Volkamer, Glyoxal processing by aerosol multiphase chemistry: towards a kinetic modeling framework of secondary organic aerosol formation in aqueous particles, *Atmos. Chem. Phys.*, 2010, **10**, 8219–8244.
- 99 A. Tilgner, P. Bräuer, R. Wolke, and H. Herrmann, Modelling multiphase chemistry in deliquescent aerosols and clouds using CAPRAM3.0i, *J. Atmos. Chem.*, 2013, **70**, 221–256.
- 100 P. Bräuer, A. Tilgner, R. Wolke, and H. Herrmann, Mechanism development and modelling of tropospheric multiphase halogen chemistry: The CAPRAM Halogen Module 2.0 (HM2), *J. Atmos. Chem.*, 2013, **70**, 19–52.
- 101 B. Ervens, A. Sorooshian, Y. B. Lim, and B. J. Turpin, Key parameters controlling OH-initiated formation of secondary organic aerosol in the aqueous phase (aqSOA), *J. Geophys. Res.-Atmos.*, 2014, **119**, 3997–4016.
- 102 C. Anastasio and J. T. Newberg, Sources and sinks of hydroxyl radical in sea-salt particles, *J. Geophys. Res.-Atmos.*, 2007, **112** (D10), D10306.
- 103 T. Arakaki, C. Anastasio, Y. Kuroki, H. Nakajima, K. Okada, Y. Kotani, D. Handa, S. Azechi, T. Kimura, A. Tsuchi, and Y. Miyagi, A general scavenging rate constant for reaction of hydroxyl radical with organic carbon in atmospheric waters, *Environ. Sci. Technol.*, 2013, **47** (15), 8196–8203.
- 104 W. W. Hu, B. B. Palm, D. A. Day, P. Campuzano-Jost, J. E. Krechmer, Z. Peng, S. S. de Sa, S. T. Martin, M. L. Alexander, K. Baumann, L. Hacker, A. Kiendler-Scharr, A. R. Koss, J. A. de Gouw, A. H. Goldstein, R. Seco, S. J. Sjostedt, J. H. Park, A. B. Guenther, S. Kim, F. Canonaco, A. S. H. Prevot, W. H. Brune, and J. L. Jimenez, Volatility and lifetime against OH heterogeneous reaction of ambient isoprene-epoxydiol-derived secondary organic aerosol (IEPOX-SOA), *Atmos. Chem. Phys.*, 2016, **16**, 11563–11580.
- 105 H. K. Lam, K. C. Kwong, H. Y. Poon, J. F. Davies, Z. Zhang, A. Gold, J. D. Surratt, and M. N. Chan, Heterogeneous OH oxidation of isoprene-epoxydiol-derived organosulfates: kinetics, chemistry and formation of inorganic sulfate, *Atmos. Chem. Phys.*, 2019, **19**, 2433–2440.
- 106 Y. Chen, Y. Zhang, A. T. Lambe, R. Xu, Z. Lei, N. E. Olson, Z. Zhang, T. Szalkowski, T. Cui, W. Vizuete, A. Gold, B. J. Turpin, A. P. Ault, M. N. Chan, and J. D. Surratt, Heterogeneous hydroxyl radical oxidation of isoprene-epoxydiol-derived methyltetrol sulfates: plausible formation mechanisms of previously unexplained organosulfates in ambient fine aerosols, *Environ. Sci. Technol. Lett.*, 2020, **7** (7), 460–468.
- 107 Y. Huang, K. M. Barraza, C. M. Kenseth, R. Zhao, C. Want, J. L. Beauchamp, and J. H. Seinfeld, Probing the OH Oxidation of Pinonic Acid at the Air–Water Interface Using Field-Induced Droplet Ionization Mass Spectrometry (FIDI-MS). *J. Phys. Chem. A.*, 2018, **122** (31), 6445–6456.
- 108 F. A. Houle, A. A. Wiegel, and K. R. Wilson, Predicting aerosol reactivity across scales: from the laboratory to the atmosphere. *Environ. Sci. Technol.*, 2018, **52** (23), 13774–13781.
- 109 F. A. Houle, W. D. Hinsberg, and K. R. Wilson, Oxidation of a model alkane aerosol by OH radical: the emergent nature of reactive uptake, *Phys. Chem. Chem. Phys.*, 2015, **17**, 4412–4423.
- 110 M. M. Chim, C. Y. Chow, J. F. Davies, and M. N. Chan, Effects of Relative Humidity and Particle Phase Water on the Heterogeneous OH Oxidation of 2-Methylglutaric Acid Aqueous Droplets. *J. Phys. Chem. A.*, 2017, **121** (8), 1666–1674.
- 111 H. Guo, L. Xu, A. Bougiatioti, K. M. Cerully, S. L. Capps, J. R. Hite Jr., A. G. Carlton, S.-H. Lee, M. H. Bergin, N. L. Ng, A. Nenes, and R. J. Weber, Fine-particle water and pH in the southeastern United States, *Atmos. Chem. Phys.*, 2015, **15**, 5211–5228.